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How urban green management is influencing passerine birds' nesting in the Mediterranean: a case study in a Catalan city

Abstract

The vegetation within the urban system provides sheltering and food provisions to birds, influencing their nesting options. This study analyses for the first time in the Mediterranean area how different socio-ecological factors related with public urban green management can influence the nesting of the passerine bird order. It uses a case study in the city of Valls (Catalonia, Spain). First, the public urban green was quantitatively and qualitatively characterised; then the nests from the passerine birds were collected and identified, and finally, potential associations between nests and urban green-related socio-ecological factors such as vegetation type (tree, shrub, herb, liana), plant species, neighbourhood type, pruning type, fruit and seed production, and presence of insect plague were analysed. A total of 300 nests were identified and belonged, mostly, to the family of Fringillidae and Sylviidae, all from Mediterranean agroforestry areas. Passerine birds show preference for the historic centre, being this area the one with highest biodiversity of vegetation in the city, in detriment of surrounding neighbourhoods, which in turn are less biodiverse. Passerine birds do not consider four tree species (Celtis australis, Laurus nobilis, Robinia pseudoacacia and Pinus pinea) suitable for nesting whereas showing preference for two tree species of medium height and size (Hibiscus syriacus and Melia azederach). Also, passerine birds seem to preferably nest in trees that have been pruned intensively. These results suggest that, to strengthen the passerine bird diversity in cities, urban green management should promote certain species of trees of medium size and intensive pruning while supporting the overall biodiversity of the urban green. All these results contribute to inform effective urban planning and management strategies for passerine birds conservation that aim to reconcile urban development and urban biodiversity protection.

Key words: Urban biodiversity, ornithofauna.

Competing interests:

Highlights:

Links between passerine birds' nesting and urban green management are tested. Intensive tree pruning positively influence passerine birds' nesting.

Passerine prefer nesting in areas with dense tree canopies and less predators.

Introduction

The richness and diversity of bird communities in cities depends on the richness and diversity of the urban green spaces. The bioclimatic area and the type and degree of urbanisation (Clergeau *et al.*, 2006) determine bird communities distribution in urban green spaces. However, the maximum richness and diversity of such bird communities is not necessarily achieved in less urbanised areas (Jokimäki and Suhonen, 1993; Carbó-Ramírez and Zuria, 2011). Bird communities select habitats of different degree of urbanisation accordingly to their habits. For instance, in high urbanised areas there are anthropophilic species, which take profit from the human activities, whereas in low urbanised areas bird species living in agroforestry vegetation turn up (Boada and Capdevila, 2000; Burger *et al.*, 2004; Marzluff and Rodewald, 2008; Parker and Nilon, 2012).

Therefore, birds can find suitable habitats under optimal conditions for their living in urban green spaces, such as appropriate microclimate and refuge, large quantities of food resources, less competition between species and less predation in the nesting areas (Ortega and MacGregor, 2009; Camprodon and Guixé, 2012). The design and management of urban green spaces will thus affect the diversity and richness of these bird communities. Two factors play a key role in this regard: the composition and the structure of the vegetation of these urban green spaces (MacGregor-Fors and Schondube, 2011). On the one hand, the composition of plant communities is intimately related to the diversity of birds (James and Wamer, 1982; Huang et al., 2015). For instance, in the city of Vinnytsia (Ukraine), researchers found significant correlations between the heterogeneity and abundance of trees with the richness and density of birds. This study also showed a positive correlation between bird diversity and plant flowering richness (Blinkova and Shupova, 2017). On the other hand, the volume and density of plants in the urban green are positively related to birds' richness and diversity (Savard et al., 2000; Mella and Loutit, 2007), so the thinning of trees and shrubs is counterproductive (Camprodon and Brotons, 2006; Yang et al., 2015). Similarly, inappropriate structure of the vegetation in green urban spaces could cause a further reduction in the diversity of birds (Ge et al., 2005; Xu et al., 2007; Yang, et al., 2015).

Particularly, trees are considered as one of the most important elements to increase bird richness and diversity in urban green spaces (Palomino and Carrascal, 2006; Yang et al., 2015; Weaving et al., 2016). Tree canopies provide sheltering, nesting sites and feeding opportunities (Munyenyembe et al., 1989; Steele and Koprowski, 2001). Specially, birds use dense tree canopies, tree trunk with holes and branches that produce fruits or seeds. It is also important to consider that the presence of these resources for birds' refuge, nesting and breeding promotes the access of adjacent flora and fauna into the urban green spaces (Briz, 1999 and 2004; Boada and Sànchez, 2012). To sum up, the promotion of urban green management actions leading to a suitable composition and structure of the vegetation in cities can potentially entail an improvement in the diversity and richness of birds living in there (Camprodon and Brotons, 2006; Shanahan et al., 2011).

The vast majority of studies about richness and diversity of birds in urban environments focus on the breeding success, e.g. by analysing nests' depredation rates and showing that it is higher when associated with specific mammals and predatory birds (Miller et al., 1998; Matthews et al., 1999; Jokimäki and Huhta, 2000; Reale and Blair, 2005; Phillips et al. 2005; Bakermans and Rodewald, 2006; Burhans and Thompson, 2006; Smith-Castro, 2008). In the case of mediterrenean cities, nest depredation is caused mainly by the presence of cats (Stracey, 2011) and magpies (Bonnington et al., 2015), though the only magpie species

present in the city (*Pica pica*) prefers the nearby rural areas instead of the urban ones (Andrén, 1992). Similarly, other works performed in peri-urban areas show that the rate of depredation of nests at low height may be higher due to the high influx of domestic animals (Miller *et al.*, 1998) whereas those located at higher height remain better conserved (Smith-Castro, 2008). However, only few research has addressed the effects of the vegetation patterns on bird nesting in urban green spaces.

The main goal of this article is to study the effect of different socio-ecological factors related to public urban green management on the richness and diversity of nests of the passerine bird order in the Catalan Mediterranean city of Valls. First, we made inventories of the ornamental vegetation in public urban green spaces of the city of Valls and characterised its biodiversity; second, passerine nests were collected, identified and characterized in the study area; and third, the relationship between both concepts was analysed. The relationship between socioecological factors and the bird nesting may become a tool for urban green managers and technicians that positively consider urban biodiversity.

Methodology

Study area: the urban area of the city of Valls.

The city of Valls is located at the northeast of the Iberian Peninsula, at an altitude of 215 m a.s.l.. The total urban area of the municipality is 2.06 km² (though the municipality contains up to 53.2 km² of non-urban spaces) and the city has a population of 22,537 inhabitants in 2018 (Ajuntament de Valls, 2018). The city has a typical Mediterranean climate characterised by soft winters and dry and warm summers. The annual precipitation is 524 mm and the average temperature is 16 °C (on average between 1993 and 2017; CMAC, 2018). The urban area is divided by three water streams that structure six residential areas which cover four different neighbourhood typologies: single-family houses, blocks of flats, isolated houses and historic centre (Table 1 and Figure 1). As expected the highest population density is found in the blocks of flats and historic centre (> 10,000 inhabitants/km²) whereas this falls significantly in the neighbourhoods of single-family houses and isolated houses (< 8,000 inhabitants/km²).

Study system: ornamental vegetation of the public urban green spaces

Between 2013 and 2015 we collected qualitative and quantitative data on the total ornamental vegetation in the urban green of Valls and produced an inventory; see Table S1 for the full inventory of the ornamental vegetation. The city has an urban green area of 110,681 m² (5.6 % of total urban area), calculated as the aerial canopy occupied by each plant individual of the ornamental vegetation (Table 1). Most of the urban green was found in the neighbourhood of single-family houses, which is also the neighbourhood type with the highest percentage of urban green spaces with respect to the total urban area. The amount of urban green area per inhabitant is 4.7 m², slightly smaller than Barcelona's green area (6.8 m²/inhab, Parcs i jardins, Ajuntament de Barcelona, 2009).

Each plant was classified as tree, shrub, herb, liana and palm. Also the following vegetation indexes were calculated in order to characterise the habitat: species richness (S), Shannon-Weaver's diversity index (H), Simpson's diversity index (D) and Pielou's evenness index (J) (Shannon, 1949; Simpson, 1949; Pielou, 1969).

Collection, storage, identification and characterisation of the passerine bird nests

We collected passerine bird nests in the urban green of Valls on January 2013 and produced the corresponding inventory in collaboration with the Park and Gardens Service of the city of Valls. Nests were kept in a dry and fresh place during 7 to 10 days for their identification (see identification criteria in Table 2). Then, they were introduced in plastic zip-lock bags together with a naphthalene ball to avoid organisms' proliferation (González, 2012). All nests come from six passerine birds: *Carduelis carduelis, Serinus serinus, Chloris chloris, Sylvia melanocephala, Turdus merula* and *Sylvia atricapilla*. We did not find other nests from different bird species in the studied urban green spaces of the city. Apart from the bird species name, we noted down the nest height and plant individual identifier, see Table 3.

Total vegetation, potential nesting vegetation and actual nesting vegetation.

As seen, the *total vegetation* is formed of all plant individuals of trees, shrubs, herbs, lianas and palms located at Valls' urban green spaces. However, not all plant species are chosen by passerine birds for nesting purposes. So, we focused on a subgroup of the total vegetation that we called the *potential nesting vegetation* which is composed of all individuals of plant species that would be hosting at least a single nest. But in order to strengthen the robustness of the analysis, we only coded a plant species as potential nesting vegetation if any of its individuals host at least three nests. Then, each individual of the potential nesting vegetation was characterised considering the following socio-ecological factors (or categorical variables): vegetation type, plant species name, neighbourhood type where the individual was placed; fruit and seed production; presence of insect plague during bird breeding period (spring's end-summer); and pruning type (see Table 3 for more details). Regarding pruning, the intensive one has the aim of controlling the volume of the canopy whereas maintenance is only used in old trees to keep the natural shape of the tree (Drénou, 2000; Ajuntament de Barcelona, 2009). Finally, we created a subgroup of the potential nesting vegetation named the *actual nesting vegetation* that is formed of all plant individuals that actually host a single nest.

Socio-ecological factors' effects in passerine bird nesting

The statistical hypothesis of this study is the following: there is an influence of the socioecological factors on passerine nesting, so the vegetation pattern of such factors in the actual nesting vegetation will be different than the one of the potential nesting vegetation. This would indicate that the categorical variable that originates such pattern is an important factor in the passerine nesting. For example, if birds select a certain tree species, the proportion of this species among the actual nesting vegetation will be statistically higher than the proportion of the same species among the potential nesting vegetation.

Contingency tables were created to study the effect of the different socio-ecological factors to passerine bird nesting, in other words the comparison of the vegetation patterns between the actual nesting vegetation and the potential nesting vegetation. The relationship between these two categorical variables was assessed using the Person's chi square statistical test (χ 2) (p < 0.01) or the Fisher's exact test (p < 0.01) when the number of observations was inferior to 5 in any of the groups. All the descriptive and inferential statistical calculations were carried out using R statistical software (The R, 2018).

Results and discussion

Characterization of ornamental vegetation biodiversity in Valls' urban green

Our inventory show that the city of Valls presents 81 different plant families, 152 genera and 239 different species (species richness, S). For the sake of a better understanding, the biodiversity of Valls is compared to Barcelona's. Though remarkable, Barcelona reaches a higher number of vegetation species, up to 1,172. All species living in the city of Valls can be found as well in the city of Barcelona. Some of the most common species from both cities coincide in the ranking of most populated species, 7 of the 15 firsts species of trees and 8 of the first 15 species of shrubs in Barcelona coincide with those of the city of Valls. Both Valls and Barcelona show a great diversity of species from all over the world, mainly from Asia, America, Africa, Oceania and as well from the rest of Europe (Argimon, 2009). So, the estimated Shannon-Weaver's diversity index (H) for Valls is 3.36 and 2.96 for Barcelona (Burriel et al., 2006). This index usually varies from 1.5 (low diversity) to 3.5 (high diversity) (MacDonald, 2003). In line with this index, the Simpson's diversity index (D) for Valls is as well high with a value of 0.88. Regarding species evenness, the Pielou's index (J) for Valls is 0.61 and for Barcelona 0.06 (Burriel et al., 2006) indicating that Valls presents more equity and a more homogenous representation for each species than Barcelona (Table 4).

When exclusively referring to trees, the number of individuals and different species is 6,376 and 103 in Valls, respectively, whereas in Barcelona are found 235,000 individuals and 200 different species (Parcs i jardins, Ajuntament de Barcelona, 2005). In Valls, though, they only represent the 12.1% of the plant individuals and 42 % of the total different species, they occupy up to 77.5 % of the total urban green area. The number of trees per inhabitant in Valls reaches a value of 0.28 whereas Barcelona shows a value of 0.15 trees/inhabitants.

The value of the Shannon-Weaver's index (H) for trees and shrubs in Valls is the highest among vegetation types, both values are above 3 and indicate a high biodiversity. However, the value for the herbs, lianas and palm species is lower than 2.0 indicating the contrary, low biodiversity in those types. Lianas and herbs species differ from trees and shrubs in the fact that they have a higher number of individuals (high richness) but they have less number of species (low diversity) (Table 4).

Characterization of nests in the urban green of Valls

We found six passerine bird species that nest in the urban green spaces of the city, which belong to three bird families, i.e., Fringilidae, Silvidae and Turdidae. These are representative of the forest ecosystems (*Quercus ilex* and *Pinus halepensis*), maquis (*Quercus coccifera*), as well as agricultural, riparian and agroforestry habitats of the Mediterranean climate regions (see Table 5 for more details) (Svensson, L., 2014; Cama, A. com. verb., 2015; Ornitho, 2015) and therefore belong to the list of possible breeding birds (Ornitho, 2015; Svensson, 2014; Cama and Filella com. verb., 2015). All of these bird families can be found in the Iberian Peninsula all year long and are well adapted to the urban system (Boada and Capdevila, 2000; Burger *et al.*, 2004; Parker and Nilon, 2012). There are other passerine breeding birds from the abovementioned ecosystems that do not nest in the vegetation of the urban green spaces but in holes or just migrate during the breeding season (see last column of Table 4), consequently these birds are not included in the study.

In total, 300 nests were collected with a population distribution clearly dominated by nests from the Fringilidae family (90.3 %) followed by Silviidae family (6.4 %), see Table 6. More than 70 % of nests come from two bird species, *Carduelis carduelis* (126 nests) and *Serinus serinus* (91 nests). It was not possible to identify the species of six nests but in all cases they belonged

to one of the already identified families of passerine birds. In the case of *Carduelis carduelis*, most of the nests (75.0 %) were found in the neighbourhoods (3a, 5a, 5b and 5c in Figure 1) close to the adjacent agroforestry area. Most of nests have been found in street trees (80.0 %) followed by the nests found in public parks (20.0 %). Indeed, according to, the streets with trees connecting urban green spaces positively influence the bird species richness, contributing with feeding and nesting sites (Fernández-Juricic et al. 2001).

Characterization of the potential nesting vegetation and the actual nesting vegetation

Among the 239 different species and 52,869 individuals of the total vegetation our data show that only 40 different species (7,798 individuals) host at least a single nest. These species would form the potential nesting vegetation. Such vegetation is still composed of trees (30 species), shrubs (8 species) and even 2 species of liana. However, if one considers the plant species that host at least three nests then the number of species that form the potential nesting vegetation is limited to 12 different species and 2,323 individuals, all of them trees (see Table 7 and Table 8 for the full list of the tree species that form the potential nesting vegetation). Therefore, the difference between the total vegetation and the potential nesting vegetation is remarkable, since the potential nesting vegetation represents only a fraction (4.4 %) of the total vegetation individuals and only trees are selected for nesting by passerine birds. These findings can be explained by the fact that trees are one of the main vegetation elements used to increase bird species richness in urban green spaces (Palomino and Carrascal, 2006; Yang et al., 2015) as tree canopies can provide sites for sheltering, breeding and feeding (Munyenyembe et al., 1989; Steele and Koprowski, 2001). In turn, the actual nesting vegetation, composed of all plant individuals of the potential nesting vegetation that host a nest, includes 267 plant individuals and their corresponding nests. Therefore, our results show that only the 11.5 % of the potential nesting vegetation is used by passerine birds. All vegetation individuals that form the potential nesting vegetation and the actual nesting vegetation are listed in Table S2.

It is also remarkable that some common tree species (> 100 individuals) of the total vegetation do not host any single nest, which would indicate that these plant species are perceived as non-adequate for nesting purposes by passerine birds, such as *Celtis australis*, *Laurus nobilis*, *Robinia pseudoacacia* and *Pinus pinea*, see Table 8.

Plant species affecting nesting

Among the potential nesting vegetation, passerine birds prefer some trees over the others. The 267 nests of the actual nesting vegetation are distributed as follows: 32.8% in *Melia azedarach*, 16.3% in *Acer negundo*, 15.5% in *Platanus hispanica*, 7.8% in *Morus alba*, 6.6 % in *Hibiscus syriacus* and 6.1 in *Ulmus minor* and 5.5% in *Sophora japonica* and 14.9 % in other species. However, the pattern for the potential nesting vegetation presents a slightly different distribution: 40.1% in *Melia azedarach*, 16.1% in *Acer negundo*, 7.5% in *Platanus hispanica*, 6.0% in *Mourus alba*, 17.6 % in *Hibiscus syracus* and 1.5 in *Ulmus minor* and 4.5% in *Sophora japonica* and 6.7 % in other species. Indeed, the Fisher's exact test states that the differences between both distributions are statistically significant (p-value = 0.0005). Therefore, birds show preference for *Hibiscus syriacus* (+11.0%) followed by *Melia azedarach* (+7.3%) and birds do not select *Platanus hispanica* (-8.0%) and *Ulmus minor* (-4.6) (Table 9). Interestingly, in the case of *Hibiscus syriacus*, if the birds nested randomly only 6.6 % of nests would be found in this species but significantly the propensity to nest in this species is three times higher, up to 17.6 %. On the contrary, *Platanus hispanica* represents 15.5% of the urban green trees, but

only 7.5 % of nests are found in this tree species (Table 9). These results confirm that bird nesting and diversity significantly relates with plant communities in the urban green (James and Wamer, 1982; Huang, et al., 2015). The different pattern of the actual nesting vegetation and the potential nesting vegetation according to the selected tree species can also be visualised in Figure 2.

Among the most common nesting birds we found in Valls' urban green, only *Serinus serinus* (Fisher's test, p-value = 0.0005) shows a specific preference for some tree species whereas *Carduelis carduelis* (p-value = 0.02) and *Chloris chloris* (p-value = 0.02) show no significant differences between the actual nesting vegetation and the potential nesting vegetation, though in both cases we observed the same tendency as that seen for *Serinus serinus*. So, *Serinus serinus* tend to select *Hibiscus syriacus* (+11.0%) and *Melia azederach* (+6.0%) whereas avoids *Platanus hispanica* (-8.8%). The selected trees are medium size trees except for *Hibiscus syriacus*, which has a small size. *Melia azedarach* generally presents medium height and size.

Preferences according to neighbourhood types

Neighbourhood types defining different urban green areas affect significantly the preferences of passerine birds nesting (Fisher's test, p-value = 0.002). The differences between the pattern of the actual nesting vegetation and that of the potential nesting vegetation by neighbourhood type (Table 11 and Figure 3) show that birds preferably nest within the historic centre (+9.4 %), and avoid the neighbourhoods with single-family houses (-5.1 %) or neighbourhoods with blocks of flats (-3.4 %). The preference for the historic center can be attributed, among other factors, to the higher canopy size and leaves' density of its trees compared to those in the surrounding area (Savard *et al.*, 2000; Mella and Loutit, 2007). Also, it could be due to a lower predator pressure in the city center (Boada and Capdevila, 2000; Boada and Gómez, 2008; Parker and Nilon, 2008; Boada and Sánchez, 2012) than in peri-urban areas where the predation rates are higher (Miller, *et al.*, 1998; Smith-Castro, 2008; Marzluff and Rodewald, 2008). The vegetation in the river banks that divide the neighbourhoods, the gardens in the neighbourhoods with single-family houses and the crop lands surrounding the isolated houses need to be considered when analysing the results because they can interfere the nesting results.

The nesting preference for the historic centre might be also related to the higher biodiversity of the neighbourhood type compared to the other ones. Table 12 lists the calculation of the biodiversity indices at the level of the neighbourhood types and indeed finds a correlation between the biodiversity and the nesting tendency. The higher the biodiversity the higher the nesting tendency with the exception of isolated houses. The historic centre shows a Shannon-Weaver's index of 3.59 which is significantly higher than that of single family houses, 2.54. This relationship has already observed elsewhere (Blinkova and Shupova, 2017).

Tree structure and characteristics that influence nesting.

Height is an important factor in the moment of nesting. 84.3 % of nests were found at medium height (2-4.5m), 10.3% at high height (>4.5m) and only 5.4 % at low height (<2m). On the one hand, the species *Melia azedarach* provides such adequate height since it is a medium sized tree and consequently is selected intensively by birds. Contrarily, the high size of the *Platanus hispanica* prevents birds for nesting as seen in previous sections. On the other hand, this preference towards medium sized trees and not bushes is due to the human and feral domestic animals' presence (Matthews *et al.*, 1999; Jokimäki and Huhta, 2000). At the same

time, bird species that tend to nest at high height in the wild decrease the nesting height in the cities due to the lack of predators (Boada and Capdevila, 2000; Boada and Gómez, 2008; Parker and Nilon, 2008; Boada and Sánchez, 2012).

68.5% of nests of the actual nesting vegetation were found in trees that are pruned intensively, which contrasts with the fact that only 50.0 % of potential nesting vegetation was pruned intensively. Therefore, birds select significantly trees with intensive pruning (Chi squared, p-value = 0.001) compared to the maintenance ones (see Figure 4). Nevertheless, they only show preference for the Melia azedarach (chi squared p-value = 0.0008) and Acer negundo (chi squared, p-value = 2.5E-5) that have been lopped and do not show preference for other species with intensive pruning such as Hibiscus syriacus (chi squared p-value = 0.37) or Morus alba (chi squared p-value = 0.37). In the case of Melia azedarach and Acer negundo, it was visually observed that nests found in intensive pruned trees are located in the middle of the annual twigs. These great numbers of twigs of Melia azedarach and Acer negundo are located where the tree was pruned in winter and also generate a large foliar mass in spring. It has already proven that there is a positive relationship between bird diversity and richness and the volume of the tree canopies (Savard et al., 2000; Mella and Loutit, 2007) and, on the contrary, low density of canopies during the breeding season reduces bird diversity and richness (Camprodon and Brotons, 2006; Yang et al., 2015). In fact, fauna finds shelter in large trees and in dense vegetation (Fernández-Juricic et al., 2001). It is also important to remark that the relationship with the intensive pruning is only found for two species, Melia azedarach and Acer negundo, whereas no effect is observed for the rest of potential nesting species, such as Hibiscus syriacus, Platanus hispanica, Sophora japonica and Morus alba. Therefore, intensive pruning cannot be considered a general recommendation for promoting passerine nesting.

86.9% of nests of the actual nesting vegetation were found in 9 plant species that produce edible fruits and the other 13.1% of nests are in 3 fruitless or non-edible fruit plant species whereas the potential nesting vegetation only shows a 78.2 % of plant species that produce edible fruits. Though the statistical effect size is not so big there is a significant relationship between the presence of fruits and bird nesting and consequently birds tend to select plants with edible fruits when nesting (Chi-squared p-value= 0.001). However, the similar relationship does not hold for the presence of insect plagues (chi-squared p-value = 0.109) and birds do not tend to select plant species that suffer from louse (insect plague) during spring or summer. Thus, passerine birds do not tend to nest in feeding areas, taking into account, first, that passerine birds are granivorous and insectivorous but not frugivorous and, second, that passerine birds feed almost entirely from insects during nesting period because insects are richer in proteins and consequently more appropriate for nourish their chicks (Svensson, 2014; Filella com. verb., 2015; Huang *et al.*, 2015).

Conclusions

In this article we investigated the relationship between several socio-ecological factors linked with public urban green management and their effects on the nesting of the passerine bird order in the Mediterranean city of Valls. Findings show that urban green biodiversity indices (Shannon-Weaver and Simpson) of the city of Valls fall within common range for a standard

Mediterranean city. While the biodiversity results for trees and bushes were above the average, herbs and lianas were below.

Passerine birds select historic centre of the city in detriment of the other type of neighbourhoods probably due to a reduced presence of predators and higher number of trees with denser canopies. Also, the historic centre hosts the highest plant biodiversity in the city according to the Shannon-Weaver's diversity index (H), which seems to favour passerine birds nesting. Most nests have been found on streets with old trees and mostly are from *Serinus serinus*, *Carduelis carduelis*, *Chloris chloris* and *Sylvia atricapilla*, which are present usually at Mediterranean agroforestry areas. Fringillidae family birds select mainly *Melia azedarach* and *Hibiscus syriacus*, which are medium-size trees. Regarding at the pruning type, birds prefer nesting on *Melia azedarach* pruned intensively. Moreover, the identified birds nest in those trees that develop edible fruits but do not show preference for those that are attacked by insects that turn into plagues, which implies that passerine birds feed outside the nesting areas since these birds are especially insectivorous in the breeding period.

To sum up, in order to increase birds' biodiversity in Valls, and other similar Mediterranean cities, biodiversity management plans should take into account that: (1) the increment of the biodiversity of the urban green attracts also passerine birds as seen in the case of the historic center, (2) the presence of medium-sized trees in the streets like *Melia azedarach* and *Hibiscus syriacus* favour passerine bird nesting and (3) the intensive pruning of *Melia azedarach* and *Acer negundo* also foster their nesting though no effect is found for other tree species such as *Hibiscus syriacus* or *Platanus hispanica*. Despite the study is not focused on the breeding success of the identified nests, results are a useful contribution for the knowledge of urban bird biodiversity. Moreover, the results could improve the management of public urban green areas in order to promote passerine nesting.

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References

- -Ajuntament de Valls, 2018. Padró municipal d'habitants. Valls.
- -Andrén, H., 1992. Corvid Density and Nest Predation in Relation to Forest Fragmentation: A Landscape Perspective. Ecology 73, 794-804.
- -Argimon, X., 2009. Estudi de la biodiversitat vegetal dels parcs i jardins de Barcelona. Fundació de l'enginyeria agrícola catalana. Parcs i jardins de Barcelona, Institut Municipal. Barcelona.
- -Banc de dades de biodiversitat de Catalunya (BDBC), 2015. Retrieved July 2017 http://biodiver.bio.ub.es/biocat/homepage.html
- -Bakermans M.H., Rodewald, A.D., 2006. Scale-dependent habitat use of Acadian Flycatcher (*Empidonax virescens*) in central Ohio. The Auk 123, 368-382.

- -Blinkova, O., Shupova, T., 2017. Bird Communities and Vegetation Composition in the Urban Forest Ecosystem: Correlations and Comparisons of Diversity Indices. Ekologia 36, 366–387.
- -Boada, M., Capdevila, L., 2000. Barcelona, Biodiversitat urbana. Ajuntament de Barcelona, Barcelona, pp. 254.
- -Boada, M., Gómez, J., 2008. Biodiversidad. Cuadernos de Medio Ambiente. Rubes, Barcelona, pp. 176.
- -Boada, M., Sànchez, S., 2012. Naturaleza y cultura, biodiversidad urbana. Eco-innovación para la Mejoría Ambiental de Productos y Servicios. E. Diagrama. Sao Carlos, pp.131-142.
- -Bonnington, C., Gaston, K.J., Evans, K. L., 2015. Ecological traps and behavioural adjustments of urban songbirds to fine-scale spatial variation in predator activity. Animal Conservation 18, 529-538.
- -Briz, J., 1999. Naturación Urbana. Cubiertas Ecológicas y Mejora Medioambiental, Ediciones Mundi Prensa Libros S.A., pp. 63-80.
- -Briz, J., Felipe, I., 2004. Incorporación de la naturaleza en cada rincón de la ciudad: Naturación urbana. Arquitectura del Paisaje: Construcción y Medioambiente 120, pp.12-19.
- -Burger, J., Jeitner, C., Jensen, H., Fitzgerald, M., Carlucci, S., Skukla, S.2004. Habitat use in basking Northern water (Nerodia sipedon) and Eastern garter (*Thamnophis sirtalis*) snakes in urban New Jersey. Urban Ecosystems 7, 17–27.
- -Burhans, D.E., Thompson, F.R., 2006. Songbird abundance and parasitism differ between urban and rural shrublands. Ecological Applications 16, 394-405.
- -Burriel, J.A., Ibáñez, J.J., Terradas, J., 2006. El mapa ecológico de Barcelona: Los cambios de la ciudad en las últimas tres décadas. Cuadernos Geográficos 39, 167-184.
- -Camprodon, J., Brotons, L., 2006. Effects of undergrowth clearing on the bird communities of the Northwestern Mediterranean Coppice Holm oak forests. Forest Ecology and Management 221, 72–82.
- -Camprodon, J., Guixé, D., 2012. Els espais urbans: manual de gestió d'hàbitat per la fauna vertebrada. Diputació de Barcelona. Obra Social La Caixa. Barcelona, pp.221.
- -Carbó-Ramírez, P., Zuria, I., 2011. The value of small urban greenspace for brids in a Mexican city. Landscape and Urban Planning 100, 213-222.
- -Centre Meteorològic de l'Alt Camp (CMAC), 2018. Retrieved July 2018 http://www.metacamp.net/
- -Clergeau, P., Croci, S., Jokimäki, J., Kaisanlahti-Jokimäki, M-L., Dinetti, M., 2006. Avifauna homogenisation by urbanisation: analysis at different European latitudes. Biological Conservation 127, 336–344.
- -Drénou, C., 2000. La poda de los árboles ornamentales. Del por qué al cómo. Mundi Prensa Libros S.A., pp. 264.
- -Duperat, M., 2005. Nidos y huevos. Susaeta ediciones. pp.144.
- -Espais verds i Biodiversitat. Parcs i jardins de l'Ajuntament de Barcelona, 2005. Retrieved July 2017 http://w110.bcn.cat/portal/site/MediAmbient/
- -Fernández-Juricic, E., 2001. Density-dependent habitat selection of corridors in a fragmented landscape. Ibis 143, 278–287.
- -Ge, Z.M., Wang, T.H., Shi, W.Y., Zhou, L.C., Xue, W.J., 2005. Impacts of environmental factors on the structure characteristics of avian community in Shanghai wood-lots in spring. Zoological Research 26, 17–24.
- -Huang, Y., Zhao, Y., Li, Sh., Gadow, K., 2015. The Effects of habitat area, vegetation structure and insect richness on breeding bird populations in Beijing urban parks. Urban Forestry & Urban Greening 14, 1027–1039.
- -González, J., 2012. Bolas de Naftalina. Ed. Bubok Publishing.
- -Harrison, C.J., 1991. Guía de Campo de los nidos, huevos y polluelos de las aves de España y Europa. Ediciones Omega. pp .482.

- -Institut Cartogràfic i Geològic de Catalunya (ICGC), 2018. Retrieved July 2018 http://www.icc.cat/
- -Institut d'Estadística de Catalunya (IDESCAT), 2018. Retrieved July 2018 http://www.idescat.cat/es/
- -James, F., Wamer, N., 1982. Relationships between temperate forest bird communities and vegetation structure. Ecology 63, 159-171.
- -Jokimäki, J., Huhta, E., 2000. Artificial nest predation and abundance of birds along an urban gradient. The Condor 102, pp. 838-847.
- -Jokimäki, J., Suhonen, J., 1993. Effects of urbanization on the breeding bird species richness in Finland: a biogeographical comparison. Ornis Fennica 70, 71–77.
- -MacDonald, G. M., 2003. Biogeography: Space, Time, and Life. John Wiley & Sons, Inc. New York.
- -MacGregor-Fors, I., Schondube, J.E., 2011. Gray vs. green urbanization: relative importance of urban features for urban bird communities. Basic and Applied Ecology 12, 372–381.
- -Marzluff, John M., Rodewald, Amanda D., 2008. Conserving Biodiversity in Urbanizing Areas: Nontraditional Views from a Bird's Perspective. Cities and the Environment 1, Article 6.
- -Matthews, A., Dickman, Ch., Major, R., 1999. The influence of fragment size and edge on nest predation in urban bushland. Ecography. Pattern and diversity in ecology 22, pp. 349–356.
- -Mella, J. E., Loutit, A., 2007. Ecologia comunitaria y reproductiva de aves en cerros islas y parques de Santiago. Unión de Ornitólogos de Chile. Boletín Chileno de Ornitología 13, pp. 13-27.
- -Miller, S.G., Knight, R.L., Miller, C.K., 1998. Influence of recreational trails on breeding bird communities. Ecological Applications 8, 162-169.
- -Munyenyembe, F., Harris, J., Hone, J., Nix, H., 1989. Determinants of bird populations in an urban area. Australian Journal of Ecology 14, 549–557.
- -Ornitho, Institut Català d'Ornitologia (ICO), 2015. Retrieved July 2017 http://www.ornitho.cat.
- -Ortega-Álvarez, R., MacGregor-Fors, I, 2009, Living in the big city: Effects of urban land-use on bird community structure, diversity, and composition. Landscape and Urban Planning 90, 189–195.
- -Palomino, D., Carrascal, L.M., 2006. Urban influence on birds at a regional scale: acase study with the avifauna of northern Madrid province. Landscape and Urban Planning 77, 276–290.
- -Parker, T. S., Nilon, C. H., 2008. Gray squirrel (*Sciurus carolinensis*) density, habitat suitability, and behavior in urban parks. Urban Ecosystems 11, 243–255.
- -Parker, T.S., Nilon, C.H., 2012. Urban landscape characteristics correlated with the synurbization of wildlife. Landscape and Urban Planning 106, 316–325.
- -Phillips, J., Nol, E., Burke, D., Dunford, W., 2005. Impacts of housing developments on woodthrush nesting success in hardwood forest fragments. The Condor 107, 97-106.
- -Pielou, E.C., 1969. An Introduction to Mathematical Ecology. Wiley-Interscience John Wiley & Sons, pp. 285.
- -Reale, J.A., Blair, R.B., 2005. Nesting success and life-history attributes of bird communities along an urbanization gradient, Urban Habitats 3, 1-24.
- -Rueda, S. (dir.), 2009. Plan de indicadores de Biodiversidad urbana de Vitoria-Gasteiz. Agencia de ecologia urbana de Barcelona. pp. 477.
- -Savard, J.L., Clergeau, P., Mennechez, G., 2000. Biodiversity concepts and urban ecosystems. Landscape and Urban Planning 48, 131–142.
- -Shanahan, D.F., Miller, C., Possingham, H.P., Fuller, R.A., 2011. The influence of patcharea and connectivity on avian communities in urban revegetation. Biological Conservation 144, 722–729.

- -Shannon, C.E., Weaver, W., 1949. The Mathematical Theory of Communication. University Illinois Press, Urbana, IL.
- -Simpson, E.H., 1949. Measurement of Diversity. Nature 163, 688.
- -Smith-Castro, J. 2008. Impacts of recreational trails on breeding birds in forested urban parks. Columbus, OH: The Ohio State University.
- -Steele, M. A., Koprowski, J. L., 2001. North American tree squirrels. Washington/London: Smithsonian Institution Press.
- -Stracey, C. M., 2011. Resolving the urban nest predator paradox: The role of alternative foods for nest predators. Biological Conservation 144, 1545-1552.
- -Svensson, L., 2014. Guia d'ocells. Europa i regió mediterrània. Edicions Omega. pp. 448.
- -The R Projecte for Statistical Computing (2018). Retrieved July 2018 https://www.r-project.org/
- -Weaving, M.J., White, J.G., Isaac, B., Rendall, A.R., 2016. Adaptation to urban environments promotes high reproductive success in the tawny frogmouth (*Podargus strigoides*), an endemic nocturnal bird species. Landscape and Urban Planning 150, 87–95.
- -Xu, X.J., Ge, Z.M., Pei, E.L., Shi, W.Y., Wang, Z.H., Wang, T.H., 2007. Avian diversity and its affecting factors in Shanghai Expo's site and surrounding areas. Chinese Journal of Ecology 26, 1954–1958.
- -Yang, G., Xu, J., Wang, Y., Wang, X., Pei, E., Yuan, X., Li, H., Ding, Y., Wang, Z., 2015. Evaluation of microhabitats for wild birds in a Shanghai urban area park. Urban Forestry & Urban Greening 14, 246–254.